1	Synthesis of reactive MgO from reject brine via the addition of $\frac{NH_4OH}{NH_4OH}$
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10	Abstract
11	Reactive magnesia (MgO) with a high purity and reactivity is used in several high-end
12	applications. This study reports the feasibility of synthesizing high reactivity MgO from reject
13	brine with the use of NH_4OH . The molar amount of NH_4OH was optimized at a NH_4OH
14	/Mg ²⁺ molar ratio of 6 to provide maximum magnesium oxide yield and purity. This led to the
15	synthesis of Mg(OH) ₂ with a purity of 93.5%, which was further calcined at 500 $^{\circ}$ C for 2
16	hours to produce reactive MgO with a SSA of 78.8 m^2/g . This study shed light on the
17	significant potential of reject brine in the recovery of Mg^{2+} and the synthesis of reactive MgO
18	with a wide range of potential applications.
19	
20	Keywords: MgO; Mg(OH) ₂ ; reject brine; NH ₄ OH; synthesis; microstructure

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22 **1. Introduction**

23 Magnesium oxide (MgO) is produced in different grades and finds use in several applications ranging from the pharmaceutical to the refractory industries due to its superior stability and 24 25 chemical resistance. A majority of MgO is produced via the calcination of magnesite (MgCO₃) through the dry route. The properties of the resulting MgO strongly depend on the 26 27 calcination conditions. Higher calcination temperatures and longer residence times lead to 28 increases in the size of the MgO grain, which results in the decrease of its specific surface 29 area (SSA) and reactivity (Shand, 2006; Mo et al., 2010). Most of the commercially available 30 MgO can be classified into four grades, depending on the conditions used during its 31 production, i.e. calcination temperature and residence time. Fused MgO is produced above the fusion temperature of MgO (2800 °C) and has the lowest SSA and reactivity, which is 32 33 excellent in chemical stability and moisture resistance (Wang et al., 2012). Dead-burned MgO is obtained at calcination temperatures above 1400 °C and has a very low SSA and reactivity. 34 35 It is widely used in the refractory industry as a fire-resistant and thermal insulation material. Hard-burned MgO is produced at 1000-1400 °C with a low SSA and limited reactivity. It is 36 37 mostly used as an expansive additive in concrete for shrinkage compensation (Gao et al., 38 2008; Mo et al., 2014). Light-burned (reactive or caustic-calcined) MgO is produced at much 39 lower temperatures ranging between 700 and 1000 °C and therefore retains a high SSA and 40 reactivity. Because of its high reactivity, the light-burned MgO has been used in various 41 applications as a fertilisers, catalyst, chemical absorbent and filtration medium (Kramer, 42 2001; Lee et al., 2004; Shand, 2006; Caraballo et al., 2009). Recent studies have shown that 43 reactive MgO can also be used as a cement binder by itself or along with Portland cement (PC) 44 and other supplementary cementitious materials, depending on the application (Liska et al., 45 2012a; Liska et al., 2012b; Al-Tabbaa, 2013; Unluer and Al-Tabbaa, 2013; Unluer and Al46 Tabbaa, 2014). The main advantages of reactive MgO cements over traditional PC are listed 47 as its significantly lower calcination temperatures (700-1000 vs. 1450 °C), ability to absorb 48 carbon dioxide (CO_2) in the form of stable carbonates while gaining strength, and complete 49 recyclability at the end of its lifetime.

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51 Unlike limestone which is abundant and available worldwide, large magnesite deposits are 52 mainly located in China and North Korea (Shand, 2006). According to RAM (2015), around 53 8.5 million tonnes of MgO is produced from magnesite annually, for which China is the 54 leading provider with a 49% market share. Lack of magnesite availability on a global level 55 highlights the need to identify alternative sources for the production of MgO, which also 56 suffers from low purity and reactivity due to the impurities present in the parent materials. 57 Today, a significant portion of global MgO supply is via the calcination of magnesium 58 hydroxide Mg(OH)₂ generated from magnesium-rich sources such as seawater or natural 59 brine, which contributes to about 14% of the global MgO production (Kramer, 2001). This 60 process involves the extraction of seawater/brine, pre-treatment and final processing steps. 61 Seawater intake structures are mostly utilized in large seawater desalination plants which 62 mainly extract seawater from open sea (Pankratz, 2004). Seawater/brine pre-treatment often 63 employs a pH adjuster to de-carbonate the solution at hand. This involves the addition of sulphuric acid to decrease the pH of the solution to 4. Seawater is then passed through a 64 65 desorption tower where it is aerated to remove CO₂ in case of the precipitation of CaCO₃ along with Mg(OH)₂ (Shand, 2006). After the seawater/brine is softened, it is then pumped 66 67 into an agitated reactor vessel, during which a strong base is added into the solution to raise 68 the pH to 10.5, enabling the precipitation of magnesium (Shand, 2006). Generally, calcium hydroxide (Ca(OH)₂) derived from calcined lime (CaO) or dolime (CaO·MgO) is deployed in practice. Friedrich et al. (1946) patented a simple process to precipitate MgO in the form of Mg(OH)₂ from seawater at a pH of 10.5 using a lime solution (Friedrich et al., 1946). Dolime is preferred due to its self-contained MgO content, enabling the use of only half of the usually required volume of seawater or brine while the other half is derived from dolime to produce the same amount of MgO as would be if lime was used (Al-Zahrani and Abdel-Majeed, 2007).

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The precipitation of Mg^{2+} can be performed via the addition of a range of alkali sources other 77 78 than lime (Turek and Gnot, 1995; Dave and Ghosh, 2005; El-Naas, 2011; Tran et al., 2013; 79 Khuyen Thi et al., 2016). Hydrated lime (Ca(OH)₂·2H₂O) can also be used to precipitate Mg(OH)₂ from seawater at a pH ranging between 7 and 7.5 (Dave and Ghosh, 2005). 80 81 However the introduction of calcium-based alkalis results in the formation of gypsum 82 (CaSO₄·2H₂O) along with other precipitates due to the presence of sulphate in seawater, 83 which necessitates the pre-treatment of seawater by adding CaCl₂ for desulfation. Another 84 additive utilized for the precipitation of Mg(OH)₂ from natural mine brine is sodium 85 hydroxide (NaOH). Turek and Gnot (1995) explored the effect of reaction temperature on the 86 sedimentation of the precipitants via the addition of NaOH into mine brine, which found out 87 the improved precipitation when reaction temperature decreased from 40 to 10 °C. The 88 improved sedimentation was attributed to the high viscosity of NaOH solution and brine, 89 which impeded the contact between the precipitating agent and mine brine and lowered the 90 diffusion rate, thereby improving the crystal structure of Mg(OH)₂ (Turek and Gnot, 1995). 91 Ammonia solution (NH_4OH) has also been reported to precipitate $Mg(OH)_2$ from 92 seawater/brine. The use of NH₄OH buffers the solution at a basic pH of around 10, which favours the precipitation of $Mg(OH)_2$. Unlike other alkalis such as lime or dolime, NH_4OH does not introduce additional cations which result in undesirable precipitates (e.g. CaCO₃), as impurities. Furthermore, NH_4OH can be recycled at the end of the reaction, which allows the design of a closed-system as suggested in the modified Solvay process and thereby eliminates the generation of waste (El-Naas, 2011).

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The precipitation of Mg²⁺ from synthetic solutions has also been performed by several studies 99 100 (Alvarado et al., 2000; Henrist et al., 2003; Yan et al., 2005). Henrist et al. (2003) studied the 101 influence of the chemical nature of the utilized bases (NaOH and NH₄OH), type of counter-102 ions and temperature on the morphological characteristics of Mg(OH)₂ precipitated from a 103 synthetic MgCl₂ solution. It was observed that the use of NaOH as the alkali source led to the 104 formation of Mg(OH)₂ with a globular cauliflower-like morphology, which consisted of small 105 particles with dispersed agglomerates; while the use of NH₄OH as the alkali source resulted in 106 a plate-like Mg(OH)₂ morphology. Single and circular plate-like particles were observed at 107 lower temperatures while particles had a tendency to inter-grow at 60 °C (Henrist et al., 2003). 108 Yan et al. (2005) reported the synthesis of Mg(OH)₂ nano-flowers by a simple hydrothermal 109 reaction of MgCl₂ and CO(NH₂)₂ without any additives. The pH of the magnesium chloride 110 solution was first adjusted at a range of 3.0-9.5 via the addition of HCl or NH₄OH. The solution was then sealed in an autoclave at 95-130 °C for 18 hours before collecting the 111 112 samples. It was found that the pH value and temperature during the initial reaction was 113 essential in controlling the morphology of particles. The initially irregular spherical particles 114 developed a nano-flower morphology with increasing pH and temperature. This was attributed to the stacking of excessive Mg²⁺ and OH⁻ on the existing plate-like crystal seeds 115 116 and therefore self-assembling into a flower like structure at high pH values (Yan et al., 2005).

118 Calcination of the resulting Mg(OH)₂ synthesized from seawater/brine or synthetic solutions 119 produces MgO (Friedrich et al., 1946; Shand, 2006). A number of studies have investigated 120 the properties of MgO calcined from Mg(OH)₂ (Eubank, 1951; Itatani et al., 1988; Choudhary 121 et al., 1994; Alvarado et al., 2000; Bartley et al., 2012). Eubank (1951) identified that the 122 reaction temperature and presence of impurities significantly influence the properties of MgO. 123 The calcination of magnesium compounds was reported to take place in two distinct stages 124 starting with the loss of water and CO₂ gases between 300 and 500 °C, which creates a porous 125 structure. Recrystallization or sintering takes place at higher temperatures (> 900 °C), 126 densifying the final material, whose porosity decreases during this process. The particle sizes 127 of calcined MgO increase with increasing calcination temperature, resulting in the decrease of 128 surface area and adsorptive capacity (Eubank, 1951). The impurities contained in MgO can 129 enhance the sintering process through the formation of vitreous phases. Alvarado et al. (2000) 130 characterised MgO prepared from three precursor magnesium salts and dolomite (Alvarado et 131 al., 2000). The SSA of calcined MgO from different magnesium compounds precursors were 132 found in the decreasing order: magnesium sulfate > magnesium nitrate > magnesium acetate > 133 dolomite. On the contrary, the particle size, degree of agglomeration and porosity displayed 134 the inverse sequence.

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The main difference between MgO obtained from the calcination of magnesite and synthetically from seawater/brine or any relevant solutions is the higher purity and reactivity of the latter (Jin and Al-Tabbaa, 2014). The main drawback of the production of MgO from seawater/brine is its higher energy consumption when compared to the dry route (17 vs. 5.9 GJ per tonne of MgO (Hassan, 2013)). However, this can be optimized as further studies are

141 performed on the reduction of the energy demands of MgO production from waste brine. This 142 is particularly critical for coastal regions with a limited amount of fresh water resources such 143 as Singapore, where desalination is considered as a feasible approach to meet residential and 144 industrial water demands. Desalination involves the removal of salts from saline water to 145 produce fresh water. Currently the desalinated water from two current running seawater 146 reverse-osmosis plants provides 100 million gallons water a day, which meets up to 25% of 147 Singapore's current water demand (PUB, 2015). According to the International Desalination 148 Association, the global daily production of desalinated water generated by 18,426 149 desalination plants worldwide exceeds 86.8 million cubic meters (IDA, 2015). It is estimated that an equivalent amount is generated as reject brine per m³ of desalinated water (El-Naas, 150 151 2011).

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153 Reject brine, which is of particular interest in this study, is a concentrated by-product with a 154 high salt concentration obtained from treating brackish water or seawater in desalination 155 plants (Adham et al., 2013). The most common way to dispose reject brine is through its 156 discharge back to the sea. However, reject brine is denser than the feedstock supply due to its 157 high salt concentration and salinity and therefore tends to accumulate at the bottom of the sea 158 when discharged through an outfall without sufficient mixing. Discharge of untreated reject 159 brine has an adverse effect on the ecosystem as it alters the flora and fauna through increased 160 salinity and directly or indirectly damages all living organisms within that particular ecosystem (Mohamed et al., 2005). A newly proposed alternative is the re-treatment of the 161 salt within the reject brine to obtain valuable materials such as Mg^{2+} , which can serve as an 162 excellent source for the recovery of MgO. Therefore, reject brine can be converted into 163

valuable and useful solids, which provides a feasible and environmental friendly use of thiswaste material (El-Naas, 2011).

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167 Although there are many studies on the reaction of synthetic MgCl₂ solution, seawater or 168 natural brine with different alkali sources (Turek and Gnot, 1995; Henrist et al., 2003; Dave 169 and Ghosh, 2005; Yan et al., 2005), limited research has been reported on the recovery of 170 valuable metals from actual reject brine obtained from desalination plant (Ahmed et al., 2003). 171 This paper reports the feasibility of synthesizing reactive MgO from reject brine collected 172 from a local desalination plant in Singapore. A comprehensive study on the reaction kinetics 173 and the physical and chemical properties of the resulting Mg(OH)₂ and MgO was performed 174 via XRD, FESEM, TG/DTA and BET analyses.

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176 **2. Materials and Methodology**

177 **2.1. Materials**

178 In this study, NH₄OH solution with analytical grade (25.0% NH₃ content) supplied by Sigma-179 Aldrich (Singapore) was used as the alkali source to react with reject brine. The reject brine 180 was collected from the Tuaspring desalination plant in Singapore, the largest desalination 181 plant in South East Asia with a capacity of 318,500 m³ desalinated water per day. During the 182 desalination process, ultra-filtration membrane technology is used to remove the suspended 183 solids and microorganisms in the seawater intake during the pre-treatment process, which is 184 followed by a two-stage seawater reverse osmosis process. Saline feed water (seawater) is 185 passed through semi-permeable membranes to produce a low-salinity water and a very saline 186 concentrate (reject brine) as a by-product, which would normally be disposed back to the sea 187 through an outfall pipe (Hyflux, 2011).

189 Unlike seawater or natural brine, reject brine contains suspended solids through the chemicals 190 added to precipitate the colloidal particles in the seawater before running through ultra-191 filtration in the desalination process. After the collection of reject brine from the plant, it was 192 filtrated through a 45 µm membrane filter to remove suspended solids before further analysis. 193 The chemical composition of the filtrated reject brine, determined via Inductively Coupled 194 Plasma-Optical Emission Spectroscopy (ICP-OES), is summarized in Table 1. Along with a Mg^{2+} concentration of 1679 ppm, the presence of other impurities (e.g. Na⁺, K⁺ and Ca²⁺) was 195 196 observed.

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Table 1 Chemical composition of the reject brine used in this study

Element/ Concentration	Na	Mg	K	Ca	Sr	В	Si	Li	Р	Al
PPM	16124.3	1679.0	808.5	563.6	5.3	4.5	0.5	0.4	0.2	0.1

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201 **2.2. Methodology**

A pre-determined amount of NH_4OH (2.12-19.08 ml) was introduced into 200 ml of reject brine as the alkaline source to investigate the influence of NH_4OH dosage on the reaction kinetics with reject brine and the chemical and physical properties of the resulting Mg(OH)₂ precipitates. According to stoichiometry, the reaction of 2 moles of OH⁻ with 1 mole of Mg²⁺ leads to the precipitation of 1 mole of Mg(OH)₂. However, a higher dosage of NH₄OH (up to 19.08 ml) had to be utilized so that supersaturation can be reached due to its relatively weak nature (K_b=1.8×10⁻⁵) as a base. 210 Reject brine and NH₄OH solution were mixed with a magnetic stirrer at a constant stirring speed (300 rpm) and room temperature (25 °C), during which a pH/thermometer probe was 211 212 inserted into the beaker to monitor the temperature and pH of the reaction. The experiment 213 was terminated after 6 hours of reaction when the pH of the solution stabilized. The solids 214 were then separated from the liquid phase through a centrifuge. Precipitates were collected 215 and washed by ultrapure water for three times. The resulting precipitates $(Mg(OH)_2)$ were 216 oven-dried at 105 °C until they reached a constant mass. They were then ground into powder 217 form passing through a 125 µm sieve, after which their chemical and physical properties were 218 determined. The remaining precipitates were calcined at a pre-determined temperature of 219 500 °C for 2 hours in the furnace to produce reactive MgO, which was then characterized for 220 its chemical and physical properties.

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222 The pH and temperature of the reaction were monitored by using a pH/thermometer (Mettler 223 Toledo pH/Ion meter S220), which was calibrated before each experiment with a standard 224 solution set at a pH of 4, 7 and 10.21, to reveal the kinetics of the chemical reactions between reject brine and NH_4OH . Mg²⁺ and Ca²⁺ ion concentrations present in the solution after the 225 226 completion of the reaction were monitored by means of an ICP-OES (PerkinElmer Optima 227 DV2000) to determine the percentage of sequestrated ions. This provided insights on the yield of $Mg(OH)_2$ and $CaCO_3$ obtained at different levels of NH_4OH addition. The solution was 228 229 diluted 10 times and filtrated through a 45 µm membrane filter to remove suspended solids. 1 230 ml of the reject brine/ NH₄OH solution (after the reaction) was then taken using a pipette and 231 added to a 10 ml acidized solution (70% nitric acid) to fully terminate the reaction. The primary Na⁺, Mg²⁺ and Ca²⁺ solutions with concentrations of 100 ppm were used for 232 233 calibration before each of the ICP-OES measurements.

235 Several techniques including X-ray powder diffraction (XRD), field emission scanning 236 electron microscopy (FESEM), thermogravimetric and differential thermal analysis (TG/DTA) 237 and Brunauer-Emmett-Teller (BET) analyses were utilized to characterize the synthesized 238 Mg(OH)₂ and reactive MgO. In preparation for these analyses, all samples were vacuum dried 239 to constant mass, followed by grinding and sieving to achieve a particle size smaller than 125 240 μm. XRD was performed via a Bruker D8 Advance with a Cu Kα source under the operation conditions of 40 Kv and 40 mA, emitting radiation with a wavelength of 1.5405 Å, scan rate 241 242 of 0.02 °/step, and a 20 range of 5 to 70°. The morphology and microstructure of the 243 synthesized samples were studied by imaging powder surface using a JSM-7600F thermal 244 FESEM. Quantitative analysis of the phases in the synthesized samples was performed via 245 TG/DTA using a PyrisDiamond TGA 4000 operated at a heating rate of 10 °C/min under air 246 flow. The SSA of the synthesized samples was analysed through the BET analysis from 247 nitrogen adsorption-desorption isotherms using a Quadrasorb Evo automated surface area and 248 pore size analyser.

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250 **3. Results and Discussion**

251 **3.1. Reaction kinetics**

Fig. 1 shows the change in pH over time during the reaction between reject brine and NH_4OH . The rate of change of pH recorded at regular intervals revealed the kinetics of the reaction under different amounts of NH_4OH . The general trend shows a steady pH after the addition of NH_4OH , followed by a gradual decrease that stabilized at a constant level when the reaction reached equilibrium. Both the initial and final pH values increased with the increase of NH_4OH dosage as more OH⁻ was available, which increased the reaction rate and shortened the reaction time. This enabled the time needed to achieve equilibrium to reduce significantly from 240 to 25 minutes as the amount of NH_4OH increased from 2.12 to 19.08 ml, as shown in Fig. 2.

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Fig. 1. pH of the reaction of reject brine with different amounts of NH₄OH

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Fig. 2. Equilibrium time as a function of NH₄OH dosage

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268 **3.2. Recovery of Mg²⁺ and Ca²⁺ from reject brine**

The formation of Mg(OH)₂ was observed via the reaction between the Mg²⁺ in the reject brine and OH⁻ provided by NH₄OH. The addition of NH₄OH also favoured the conversion of HCO³⁻ (origin from CO₂ in air dissolved in the reject brine) to CO₃²⁻. This led to a reaction of CO₃²⁻ with Ca²⁺ and resulted in the precipitation of CaCO₃. The reaction paths observed during this process are shown in Eqns. 1 to 4 (Shand, 2006).

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$$\operatorname{NH_4OH}_{(aq)} \leftrightarrow \operatorname{NH_4^+} + \operatorname{OH^-}$$
 (1)

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$$Mg^{2+} + 2OH \rightarrow Mg(OH)_2$$
 (2)

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$$NH_4OH_{(aq)} + HCO_3^- \leftrightarrow CO_3^{2-} + NH_4^+ + H_2O$$
(3)

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$$\operatorname{Ca}^{2+} + \operatorname{CO}_3^{2-} \to \operatorname{CaCO}_3$$
 (4)

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Fig. 3 shows the molar percentage of Mg^{2+} and Ca^{2+} recovered from reject brine through the 279 280 addition of different amount of NH_4OH , which was calculated by measuring the ion concentration in the reject brine both before and after the reaction. The amount of Mg^{2+} and 281 Ca^{2+} precipitated from the reject brine increased with the amount of NH₄OH addition. The 282 283 precipitation of Mg^{2+} took place earlier than Ca^{2+} due to lower solubility product constant of $Mg(OH)_2$ (1.8×10⁻¹¹ mol³ l⁻³) than that of CaCO₃ (3.8×10⁻⁹ mol² l⁻²) (Sillén et al., 1964). In 284 addition, the ion product $[Mg^{2+}][OH^{-}]^{2}$ was calculated to be $7 \times 10^{-10} \text{ mol}^{3} \text{ }^{-3}$ at pH 10 (Fig. 1), 285 which was larger than the solubility product constant of Mg(OH)₂ $(1.8 \times 10^{-11} \text{ mol}^3 \Gamma^3)$ (Sillén 286 et al., 1964). This caused supersaturation of $Mg(OH)_2$ and enabled the reaction between OH^- 287 with Mg^{2+} to precipitate $Mg(OH)_2$. As can be seen, precipitation of CaCO₃ was hindered at 288 lower NH₄OH dosage (< 6.36 ml). This may be attributed to the lower pH of the solution, 289

which was unfavourable the conversion of HCO_3^{-1} to CO_3^{-2-1} and the formation of CaCO₃, when 290 less NH₄OH was added. Therefore, a higher content of Mg^{2+} than Ca^{2+} was precipitated when 291 292 a lower amount of NH_4OH (< 6.36 ml) was used. As the amount of NH_4OH increased beyond 6.36 ml, the increased amount of OH⁻ in the solution achieved a sufficient level to react with 293 both Mg^{2+} , CO_3^{2-} and Ca^{2+} , thereby increasing the amount of CaCO₃ in the precipitates. As 294 the amount of NH_4OH exceeded 16.96 ml, the percentage of Mg^{2+} and Ca^{2+} recovered 295 296 reached a plateau due to the equilibrium reached in the supernatant liquid.





Fig. 3. Percentage of Mg^{2+} and Ca^{2+} sequestrated from reject brine as a function of NH_4OH 299 dosage

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The Mg^{2+}/Ca^{2+} ratio was also used as an indicator of the purity level of the resulting $Mg(OH)_2$ 302 precipitates. As seen in Fig. 3, Mg²⁺/Ca²⁺ peaked at about 12.5 when 6.36 ml of NH₄OH was 303 added into the reject brine. This was because the initial pH was lower than 10.3, the formation 304 305 of carbonates was prevented, therefore reducing the supersaturation for CaCO₃ and contributing to a lower precipitation of CaCO₃ in the solution. Therefore, a NH₄OH to Mg²⁺ 306

ratio of 6 was determined as the most optimum ratio out of the different values used in this
study for the precipitation of Mg(OH)₂, which was further calcined to produce reactive MgO
with a high purity.

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311 **3.3.** Characterization of the synthesized Mg(OH)₂

312 **3.3.1. XRD**

313 Fig. 4 shows the XRD diffractograms of synthesized Mg(OH)₂ obtained from the reaction 314 between reject brine and different amounts of NH₄OH. The crystalline peaks observed were 315 mainly attributed to Mg(OH)₂ along with smaller amounts of CaCO₃, which was present in the crystal form of aragonite. The presence of Mg^{2+} in the reject brine inhibited the precipitation 316 317 of calcite and favoured the formation of aragonite, which was in line with the findings of 318 earlier studies (Berner, 1975). As shown in Fig. 4, the relative intensity of the peaks 319 corresponding to $Mg(OH)_2$ and $CaCO_3$ varied as the amount of NH_4OH increased, which 320 reflected the changes in the crystallinity and chemical composition of the precipitants. The 321 relative intensity of the peaks corresponding to CaCO₃ increased along with the increasing 322 amount of NH_4OH , which indicated the increased content percentage of CaCO₃ in the 323 precipitants.



3.3.2. FESEM

The morphologies of the synthesized Mg(OH)₂ obtained with the addition of different amounts of NH₄OH are shown in Fig. 5. A flake-like morphology with an average agglomerate size of 10-15 µm was observed in all cases, similar to the findings reported earlier in literature (Alvarado et al., 2000; Behij et al., 2013; Guo et al., 2015). As shown in the Fig. 5, the amount of NH₄OH did not have a significant influence on the morphology of Mg(OH)₂ synthesized from reject brine.





Fig. 5. FESEM images of Mg(OH)₂ obtained from the reaction of 200 ml reject brine with (a)
2.12, (b) 4.24, (c) 6.36, (d) 8.48, (e) 10.6, (f) 12.72, (g) 14.84, (h) 16.96 and (i) 19.08 ml of
NH₄OH

344 3.3.3. TG/DTA

Fig. 6 illustrates a typical TG/DTA graph of the synthesized $Mg(OH)_2$ obtained by adding 6.36 ml of NH_4OH to react with the reject brine. The dehydration of $Mg(OH)_2$ took place between 340 and 440 °C and resulted in a weight loss of around 25.5% due to the loss of H_2O from the system. The second reaction, attributed to the decarbonation of CaCO₃, which led to the release of CO₂, was observed between 650 and 750 °C, resulting in a weight loss of 2.5%.



Fig. 6. Typical TG/DTA curve of Mg(OH)₂ obtained from the reaction of 200 ml reject brine
with 6.36 ml of NH₄OH

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355 Table 2 summarizes the peak temperatures observed during the thermal decomposition of 356 Mg(OH)₂ and corresponding weight losses at each reaction. The weight loss due to the 357 dehydration of Mg(OH)₂ (i.e. between 340 and 440 °C) showed a decreasing trend, while the 358 weight loss due to the decarbonation of CaCO₃ (i.e. between 650 and 750 °C) increased with 359 the amount of NH_4OH . The peak temperature at which CaCO₃ decomposed into CaO and CO₂ 360 increased significantly from 655.4 to 747.2 °C with the increase in the amount of NH₄OH. 361 This could be attributed to the poorly crystalline structure of CaCO₃ that formed when lower 362 amounts of NH₄OH were reacted with the reject brine (Johnson, 2002). This was consistent 363 with the XRD patterns shown earlier in Fig. 4, where sharper and narrower CaCO₃ peaks were 364 detected under higher NH₄OH dosages.

Table 2 TG/DTA results of Mg(OH)₂ obtained from the reaction of 200 ml reject brine with

-	<mark>NH₄OH</mark>	1 st peak	Weight loss between	2 nd peak	Weight loss between
	(ml)	temperature (°C)	340-440°C (%)	temperature (°C)	650-750°C (%)
=	2.24	391.3	24.9	655.4	0.7
	4.24	393.4	24.2	691.0	1.3
	6.36	406.8	25.5	702.9	2.5
	8.48	393.4	25.1	691.0	1.5
	10.6	407.3	25.3	706.9	2.2
	12.72	408.9	24.5	706.9	2.9
	14.84	409.8	23.8	738.1	4.0
	16.96	403.8	23.0	725.4	4.9
	19.08	405.8	20.0	747.2	9.2

different amounts of NH₄OH

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Table 3 lists the compositions of the synthesized Mg(OH)₂, which were calculated according 369 370 to the TG/DTA results listed in Table 2 as well as the ICP-OES results shown in Fig. 3. Both 371 measurements revealed similar trends, which indicated a decrease in the amount of Mg(OH)₂ 372 accompanied with an increase in the amount of CaCO₃ as the amount of NH₄OH increased. 373 According to the result of TGA, the synthesized Mg(OH)₂ achieved up to 98% in purity when 374 2.24 ml of NH₄OH was added into the reject brine. While ICP-OES test indicated the highest purity of Mg(OH)₂ of 97.3% was achieved when 6.36 ml of NH₄OH was added. However, 375 376 low dosages of NH_4OH cannot achieve the supersaturation for the formation of $Mg(OH)_2$, leading to a lower yield of precipitates. As such, a NH₄OH to Mg²⁺ ratio of 6 at a NH₄OH 377 content of 6.36 ml, which led to a Mg^{2+} yield of 70% (Fig. 3) and a $Mg(OH)_2$ with a purity of 378 379 93.5% (Table 3), was chosen for the large scale production of Mg(OH)₂, which was then 380 calcined to produce reactive MgO.

381

382 Table 3 Chemical composition of the synthesized Mg(OH)₂ based on TG/DTA and ICP-OES

383

results

	TG/D	TA	ICP-OES			
<mark>МП4ОП</mark> (IIII)	$Mg(OH)_2(\%)$	$CaCO_3(\%)$	$Mg(OH)_2(\%)$	$CaCO_3(\%)$		
2.24	98.0	2.0	96.6	3.4		
4.24	96.4	3.6	96.9	3.1		
6.36	93.5	6.5	97.3	2.7		
8.48	95.9	4.1	90.2	9.8		
10.6	94.2	5.8	89.2	10.8		
12.72	92.4	7.6	86.9	13.1		
14.84	89.5	10.5	80.9	19.1		
16.96	87.0	13.0	78.4	21.6		
19.08	75.6	24.4	78.3	21.7		

385 3.4. Characterization of the synthesized reactive MgO

386 **3.4.1. XRD**

The Mg(OH)₂ obtained at a NH₄OH to Mg²⁺ ratio of 6 was calcined at 500 °C for 2 hours in a 387 388 furnace for the production of reactive MgO, whose XRD patterns are shown in Fig. 7. The 389 main peak positions of the synthesized reactive MgO matched well with the reference peaks 390 of MgO (JCPDS # 89-7746) with few minor peaks attributed to the presence of calcite, which 391 may have formed from the transformation of aragonite at higher temperatures (Kontoyannis 392 and Vagenas, 2000). The lack of Mg(OH)₂ peaks was an indication that all the initially used 393 brucite fully decomposed at the calcination conditions used (i.e. 500 °C for 2 hours), resulting 394 in the formation of reactive MgO.





397

Fig. 7. XRD diffractograms of the produced reactive MgO

399 **3.4.2. FESEM**

400 Figs. 8(a) and (b) illustrate the morphology of the reactive MgO produced from the 401 calcination of Mg(OH)₂ at 500 °C. A plate-like morphology, which was inherited from the 402 parent material brucite as was shown in Fig. 5, was observed throughout the microstructure of 403 MgO. Different from the closely packed structure of brucite, MgO demonstrated a more 404 porous structure due to the loss of water molecules during the decomposition of brucite into 405 MgO. This finding was in agreement with previous studies (Alvarado et al., 2000; Mo et al., 406 2010), in which the porous plate-like morphology of MgO calcined from magnesite was 407 reported. The main difference between the morphology shown in Fig. 8 and those reported in relevant literature was the much smaller particle size of the MgO grains in the former due to 408 409 the low calcination temperature (500 °C) utilized in its production.



Fig. 8. FESEM images of the reactive MgO produced at 500 °C at different magnifications of
(a) x14,000 and (b) x30,000

415 **3.4.3. BET**

416 Fig. 9 presents a comparison between the SSA values of the reactive MgO obtained from 417 reject brine in this study and 12 different commercial MgO samples presented in an earlier 418 study (Jin and Al-Tabbaa, 2014). The synthesized reactive MgO, referred to as "SRM", 419 achieved a SSA of 78.8 m²/g, which was higher than the SSA of most of the MgO samples 420 obtained via the dry route (i.e. through the calcination of magnesite). Alternatively, the SRM 421 revealed a lower SSA when compared to the MgO samples obtained via synthetic routes. This 422 could be attributed to the higher impurity level of SRM (90.8%) compared to the commercial 423 synthetic MgO samples (e.g. DSP (99.2%) and N50 (97.7%)). The presence of the main 424 impurity CaCO₃, who also possessed a low SSA, have also contributed to the generally lower 425 reactivity of SRM amongst the synthetically produced MgO samples.

426

427 Regardless of the different SSA comparisons, the SRM obtained in this study can be 428 classified under "category I" referring to highly reactive MgO with an SSA of over 60 m^2/g 429 according to the classification of reactive MgO proposed by (Jin and Al-Tabbaa, 2014).

Furthermore, it must be kept in mind that the SSA and therefore the reactivity of the MgO
synthesized from the reject brine as presented in this study can be further increased with an
optimization of production conditions.

433



434

Fig. 9. SSA of the synthesized reactive MgO (SRM) compared with 12 commercial MgO

samples reported in (Jin and Al-Tabbaa, 2014)

436

437

438 **4. Conclusions**

This study shed light on the significant potential of reject brine in the recovery of Mg^{2+} and the synthesis of reactive MgO with a wide range of potential applications. The obtained results demonstrated the feasibility of synthesizing reactive MgO from reject brine obtained from a desalination plant. NH₄OH was used as alkali source to precipitate Mg(OH)₂ from reject brine. The influence of the amount of NH₄OH was investigated to optimize the yield and increase the purity of the precipitates. An optimum NH₄OH to Mg²⁺ ratio of 6, which resulted in a high Mg content while minimizing Ca-based impurities, was determined. This led to the synthesis of $Mg(OH)_2$ with a high purity of 93.5%, which was further calcined at 500 °C for 2 hours to produce highly reactive MgO with a SSA of 78.8 m²/g. A comparison of the produced MgO with commercially available reactive MgO samples indicated its capability to be used in different applications necessitating the use of reactive MgO with a high SSA.

450

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